Sustaining Coral Reef Fisheries of Puerto Rico

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Report of Technical Workshops #1 and #2 October 2013

ROSENSTIEL SCHOOL of MARINE & ATMOSPHERIC SCIENCE

Fisheries Technical Workshop #1 and #2 "Length-Based Stock Assessment of Puerto Rico Reef Fishes & Computer-based Tools Laboratory"

Workshop Goals.- The goals of this 3-day technical workshops will be to (1) present research results on the development and application of length-based approaches to sustainability analysis of Puerto Rico reef fishes, and (2) to conduct hands-on training in the use of computer tools for length-based stock assessment and for management forecasting. On workshop Day 1, research results will be presented on data assimilation, analysis, and modeling aspects of length-based assessments of Puerto Rico reef fishes. On workshop Day 2, a laboratory will be conducted for training in the use of the MAST computer tool for carrying out length-based assessments. On workshop Day 3, participants will build up the results they found on day 2 by brainstorming management actions which could be implemented given the assessment results. They will then run simulations to investigate the potential outcomes of their management ideas using the MAST platform.

Target Audience.- This workshop is intended for reef fisheries scientists and managers from government and academia, as well as representatives from the commercial and recreational fishing communities and non-governmental organizations in Puerto Rico. This workshop will accommodate participants based in the Mayaguez regions.

Background.- The coral reef fisheries of the Puerto Rico reef ecosystem support multimilliondollar fishing and tourism industries. The sustainability of these fisheries is a key conservation concern given their economic and ecological importance, the significant dependence of subsistence and artisanal fishers on reef fisheries for their livelihoods, and the considerable and growing threats to coral reef habitats (i.e. coral bleaching and disease, pollution and climate change). Sustainability refers to the ability of an exploited stock to produce goods and services, including yields at suitable levels in the short term, while maintaining sufficient stock reproductive capacity to continue providing these goods and services into the indefinite future. The data- and model-limited situations confronting most coral reef fisheries, including those of Puerto Rico, have hampered application of modern stock assessment techniques that meet the legal mandate of the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA). We have developed a class of length-based methods for stock assessment of datalimited fisheries (Ault et al. 2008). These approaches have relatively simple data requirements, provide a community-level perspective on exploitation effects, and also enable evaluation of stock-specific sustainability that conforms to the legal requirements of the MSFCMA.

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OVERVIEW

The sustainability of coral reef fisheries of the Puerto Rico reef ecosystem is a critical ecological and economic concern as it is, inhabited by hundreds of reef fishes and macroinvertebrates, and supports multimillion-dollar fishing and tourism industries. Towards this end and with assistance from DNER personnel, length composition data were obtained from long-term fishery-dependent and fishery-independent sampling programs with focus on the two principal fishery-dependent sampling programs, the Trip Interview Program (TIP) that samples commercial catches and the Marine Recreational Fisheries Statistical Survey (MRFSS) that samples recreational catches. Unique capacity has been built within the Ault lab at the University of Miami to utilize these data sources for rigorous stock assessment when they were traditionally considered insufficient. To support this work both fundamental advances in development of analytical models for length based mortality assessment and the construction of a computational platform for fishery simulation (MAST) has been achieved. The results of this latest assessment continue to support the conclusion that the majority of species in the Puerto Rican fishery are significantly over exploited. There is evidence that organic transition is occurring within the commercial fishery to reduce effort and increase minimum capture size as seen in the TIP data. However management action is still required as a budding recreational fishery possesses the potential for rapid expansion which would negate and possibly reverse the positive trends being seen from the changing commercial fishery. Particular concern should be given to the parrotfish complex which display extreme overexploitation and the grouper complex which while virtually absent still make intermittent appearance in the catch data suggesting that their populations may have already undergone collapse.

INTRODUCTION

The coral reef fisheries of the Puerto Rico reef ecosystem, inhabited by hundreds of reef fishes and macroinvertebrates, supports multimillion-dollar fishing and tourism industries (**Fig. 1**). The sustainability of multispecies coral reef fisheries in Puerto Rico is a key conservation concern given their economic and ecological importance, the significant dependence of subsistence and artisanal fishers on reef fisheries for their livelihoods, and the considerable and growing threats to coral reef habitats (i.e. coral bleaching and disease, pollution and climate change).

Figure 1.- Map of the northern Caribbean Sea, with inset showing the island archipelago of Puerto Rico and the La Parguera region.

Sustainability refers to the ability of an exploited stock to produce goods and services, including yields at suitable levels in the short term, while maintaining sufficient stock reproductive capacity to continue providing these goods and services into the indefinite future. The data- and model-limited situations confronting most coral reef fisheries, including those of Puerto Rico, have hampered application of modern stock assessment techniques that meet the legal mandate of the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA). The objective of this research is to develop a quantitative toolbox of data assimilation and length-based fishery assessment methods to compute reference points for the multispecies coral reef fisheries that will facilitate the conservation efforts of state and federal managers and increase regional capacity to build sustainable reef fisheries in Puerto Rico. Our proposed approach is novel in that it has relatively simple data requirements and provides a community-level perspective on exploitation effects, yet also enables evaluation of stockspecific sustainability that conforms to the legal requirements of the MSFCMA. The long-term conservation outcome is to achieve sustainable levels of fishing for exploited groupers, snappers, and parrotfishes in Puerto Rico through development of new framework of

assimilation and modeling methods and by working closely with regional managers and scientists to implement the framework.

Major Threats and Opportunities

Intensive exploitation and overfishing are perhaps the major threats to this ecosystem. Overfishing has been identified as one of the top three global threats by the NOAA Coral Reef Conservation Program. Substantial reductions in commercial reef fish catches in Puerto Rico over the past several decades have been observed, resulting in harvest moratoria on several species (for example Nassau grouper and goliath grouper). At risk are the economic well-being of coastal human communities as well as the ecological resilience and sustainability of the coral reef ecosystem. Reef fisheries in Puerto Rico target important guilds of fish species including top predators (e.g., groupers and snappers) and herbivores (e.g., parrotfishes). Overfishing not only threatens the long-term viability of individual species, but also impacts the overall health of the coral reef ecosystem. For example, depletion of top predators can disrupt the ecosystem food web, and depletion of herbivores can give a competitive advantage to algae over coral for available substrate, in turn leading to loss of coral cover and general degradation of coral reefs. The end result of these impacts may be a general decline in the productive capacity of the coral reef ecosystem to produce fishery yields and an increased vulnerability to climate change.

Insufficient and poor quality data, and lack of an appropriate modeling framework have prevented sophisticated evaluations of the sustainability of reef fisheries. Generally lacking are the data needed to conduct modern stock assessments, including demographic rates and historical time-series of age-size structured catches by species, and the associated fishing effort by gear and sector. While the quality and scope of reef fishery catch-effort data have generally improved in Puerto Rico over the past two decades for the commercial fleet, comparable data from the recreational fleet are not available prior to 2000. There is an obvious lack of a quantitative framework to assess the sustainability of reef fishes in Puerto Rico and to guide management decision-making.

An opportunity to overcome these data and model limitations has recently been provided by Ault et al. (2008), who developed a new quantitative system of length-based empirical estimation and numerical model analyses to evaluate exploitation status via resource reference points (or sustainability benchmarks) for coral reef fishes of the snapper-grouper complex in Puerto Rico. Of the 25 reef fish species assessed, 16 were overfished according to federal benchmarks for sustainability, six were fished sustainably, and three could not be reliably determined due to insufficient data. These findings indicate that a majority of snapper-grouper species in Puerto Rico are currently fished at unsustainable levels.

Over the past decade, requirements for fishery assessment and management in the USA have moved towards a more precautionary approach that strives to 'prevent overfishing while achieving, on a continuing basis, the optimal yield from each fishery for the United States fishing industry' (MSFCMA [Magnuson-Stevens Fishery Conservation and Management Act]. Under this legal framework, determination of the sustainability of a fishery must also consider relevant socioeconomic and ecological factors, particularly whether fishing could deleteriously impact the reproductive capacity of the resource. This new process involves regulation of fishing mortality rate, over which management has some direct control, and how it should change depending upon stock reproductive potential and associated fishery yields. The data limited situations confronting coral reef fisheries in Puerto Rico have hampered application of modern stock assessment techniques that meet the legal mandate of the MSFCMA.

The use of length composition data and associated models to assess Caribbean reef fisheries is relatively novel, and the analysis of Ault et al. (2008) represented a first attempt to apply these assessment methods across the exploited reef fish community in Puerto Rico. There are some advantages to this length-based assessment method for estimating total mortality because it has relatively simple data requirements and has been shown to have relatively robust properties for assessing exploitation impacts on coral reef fishes. Unique owing to its zero-bias properties at equilibrium, the method is also relatively insensitive to trends in recruitment and exhibits desirable properties for detecting statistical differences between sustainable and nonsustainable rates (Ehrhardt and Ault 1992; Ault et al. 2005). It also enables evaluation of stockspecific sustainability that conforms to the legal requirements of the MSFCMA.

1.0 DATA SOURCES AND PROCESSING

Length composition data were obtained from long-term fishery-dependent and fisheryindependent sampling programs with the assistance of DNER personnel (**Table 1.1**). Initial analyses focused on the two principal fishery-dependent sampling programs, the Trip Interview Program (TIP) that samples commercial catches and the Marine Recreational Fisheries Statistical Survey (MRFSS) that samples recreational catches. Data processing procedures were developed using SAS statistical software to create analysis-ready length-composition datasets for both the TIP and MRFSS databases, including creating uniform data codes for species, time, space, gears, etc., to facilitate comparative analyses among data sources (**Fig. 1.1**) with annual sample size by species included for TIP (**Table 1.2**) and MRFSS (**Table 1.3**). Maps showing bathymetric and habitat features of the coastal region of Puerto Rico along with the numbering scheme for coastal municipios are provided in **Figs. 1.2A-B**.

Table 1.1: Data sources analyzed for species length composition.

Table 1.3: Observation sample size by species and year for MRFSS dataset (2000-2010). Green highlighted species have sufficient length frequency data and life history data for exploitation rate analysis. Blue highlighted species have sufficient length frequency data but are lacking life history parameters.

Figure 1.2: A. Bathymetry map showing the 35m contour line, **B.** Map of Puerto Rico showing the municipios and benthic habitat types.

Figure 1.3: Map of Puerto Rico showing the four fishing regions and corresponding principal municipios for the different databases.

2.0 LBAR ESTIMATION METHODOLOGY

Estimation of Lbar, the average length in the exploited phase, requires designation of the minimum capture length, Lc, above which fishes are generally retained in the catch and brought back to the dock for sale, consumption, etc. In cases where fishery regulations specify a minimum legal size, Lc is set to this size. In cases where a minimum size is not regulated, Lc must be estimated from the catch data. **Figures 2.1-2.7** illustrate sample frequency distributions with Lc, Lbar, and Linf/Lmax for seven focus species: red hind, dog snapper, schoolmaster, bluestriped grunt, lane snapper, redtail parrotfish, and stoplight parrotfish. The values of Lc were calculated for all years of data, while the length frequency represents just the last three years (2008-2010) being the current status.

Exploited phase length composition data were extracted from the two main fisheryindependent sampling programs, the SEAMAP program utilizing trap and bottom line gears and the NOAA Biogeography Team diver visual survey, for the reef species present in the TIP and MRFSS catch sampling. Both fishery-independent sampling programs were predominately conducted in the West-Southwest region (**Fig. 2.8**). Comparability of Lbar estimates among the various fishery-dependent and -independent sampling programs was evaluated.

Estimates of average length *L* were computed using the general ratio-of-means estimator

$$
\hat{R} = \frac{\overline{y}}{\overline{x}} = \frac{\sum_{i} y_i}{\sum_{i} x_i} , \quad (2.1)
$$

where

 $\hat{R} = \overline{L}$

 x_i = number of fish measured in trip *i*,

yⁱ = summed length of all fish measured in trip *i*.

 $n =$ number of trips,

$$
\overline{x} = \frac{\sum_{i} x_i}{n}
$$
 = sample mean of number of fish measured, and
\n
$$
\sum_{i} y_i
$$

n $\overline{y} = \frac{\sum_i y_i}{\sum_i y_i}$ $=\frac{1}{1}$ = sample mean of summed length of measured fish.

The ratio estimate \hat{R} (i.e., the estimate of average length) was then used to compute the sample variance

$$
s^{2} = \frac{(y_{i} - \hat{R}x_{i})^{2}}{n-1}, \quad (2.2)
$$

variance of the estimate,

$$
\text{var}(\hat{R}) = \frac{s^2}{n\bar{x}^2} \quad , \qquad (2.3)
$$

and standard error

$$
SE(\hat{R}) = \sqrt{\text{var}(\hat{R})} \quad . \quad (2.4)
$$

Illustrative results of these analyses are shown in **Figures 2.1-2.7**. Red hind, lane snapper, and stoplight parrotfish Lbar estimates for the three consecutive years were comparable for the TIP and MRFSS catch sampling, and indicated no directional bias in Lbar by data source (**Figures 2.1, 2.5, 2.7**). In contrast, Lbar estimates for dog snapper were higher in TIP catch sampling compared to MRFSS catch sampling (**Figure 2.2**) suggesting a difference in selectivity patterns between the two fleets, likely due to variation in the primary gears used.

Our analyses suggest that the commercial and sport fleets in Puerto Rico are sampling the same population of a given reef fish species, and that slight differences in length composition between the two sampling programs are mostly a function of how different gears are deployed in different regions and depth ranges.

Annual estimates of Lbar based solely on the commercial fleet data, the largest and longest source of length composition data, were nearly identical to Lbar estimates using the combined data from all sources and as such were used as representative for the time-series estimates of Lbar. These were computed for all reef-fish species using the following procedure for cases with small sample sizes: (i) data were pooled for up to four consecutive years if they each contained less than 15 observations from less than 5 trips until their aggregate exceeded these thresholds; (ii) individual years with poor data were retained as independent estimates with high uncertainty, (iii) If larger blocks of time still possessed insufficient sampling the species was not considered for full time-series analysis, (iv) current status estimates were still computed if possible. Species sample sizes by year are provided for TIP (**Table 1.2**) and MRFSS (**Table 1.3**) data. Time-series estimates of Lbar for 28 reef-fish species are given in **Table 2.1**. The current status (2008-2010) Lbar estimates for 34 reef-fish species are given in **Table 2.2**.

species Common Lc 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 *Cephalopholis fulvus* coney 220 247.66 248.90 245.99 250.51 250.80 252.68 256.36 255.84 256.14 254.44 256.07 256.83 *Epinephelus guttatus* red hind 260 320.84 326.20 331.55 334.24 323.61 326.43 327.82 320.81 321.60 325.22 324.80 345.99 *Lutjanus buccanella* blackfin snapper 220 270.52 302.20 282.33 280.14 260.59 281.22 274.24 275.96 275.96 275.96 288.50 316.67 Lutjanus jocu dog snapper 250 339.87 416.72 393.61 374.16 391.78 369.26 386.75 327.00 429.55 429.55 429.55 396.53 Lachnolaimus maximus hogfish 240 357.15 386.89 389.31 354.95 374.23 365.38 349.83 365.79 337.19 360.54 334.94 352.80 Lutjanus synagris lane snapper 210 248.91 251.77 247.45 249.39 250.43 254.70 246.36 265.15 251.53 244.41 250.98 245.65 Lutjanus mahogani mahogany snapper 250 296.78 293.19 293.19 286.62 282.18 292.45 292.60 288.03 296.87 296.87 295.85 304.60 Lutjanus analis mutton snapper 240 363.38 392.47 374.79 420.47 417.27 468.57 389.68 426.39 398.19 349.08 457.58 335.74 *Etelis oculatus* queen snapper 290 508.29 408.57 438.35 371.49 403.85 405.93 376.90 376.90 376.90 376.90 487.63 467.56 Lutjanus apodus schoolmaster 220 293.29 295.70 304.76 293.85 303.29 301.60 322.48 324.39 320.24 350.41 278.36 278.39 *Lutjanus vivanus* silk snapper 220 274.93 307.16 282.84 278.06 268.67 280.97 273.50 277.95 272.46 279.64 279.10 290.38 Rhomboplites aurorubens vermilion snapper 200 221.42 232.16 223.92 219.71 219.63 218.94 220.19 220.35 229.00 236.31 221.50 255.43 *Pristipomoides aquilonaris* wenchman 180 212.76 312.13 276.69 261.27 255.15 226.45 226.45 278.89 278.89 278.89 278.89 *Ocyurus chrysurus* yellowtail snapper 240 291.07 291.91 294.96 298.41 293.03 290.17 289.36 306.29 289.84 302.19 322.04 305.75 *Haemulon sciurus* bluestriped grunt 200 239.95 233.63 234.67 238.10 236.67 247.96 247.38 242.54 241.49 235.41 241.22 236.71 *Anisotremus virginicus* porkfish 210 239.45 234.24 234.00 248.36 241.11 235.77 236.70 236.70 235.59 235.59 240.07 235.81 *Haemulon plumieri* white grunt 200 233.27 229.53 232.68 234.19 235.25 234.03 237.00 238.61 230.59 238.98 231.52 231.84 *Acanthostracion polygonia* honeycomb cowfish 200 230.92 247.24 239.12 244.83 266.71 240.74 241.74 263.86 263.86 263.86 261.60 257.56 *Balistes vetula* queen triggerfish 230 292.77 284.74 282.53 292.81 294.86 296.92 295.00 272.65 296.13 274.25 299.88 343.85 *Acanthostracion quadricornis* scrawled cowfish 190 227.42 228.20 238.25 218.05 231.04 229.63 241.08 231.50 242.11 242.11 228.93 238.94 *Lactophrys triqueter* smooth trunkfish 150 192.86 204.89 196.05 189.36 195.00 197.46 214.82 213.50 213.50 213.50 185.71 234.83 *Pseudupeneus maculatus* spotted goatfish 150 190.14 186.16 182.96 193.82 195.17 190.44 173.98 188.45 198.00 179.58 167.90 167.11 *Lactophrys bicaudalis* spotted trunkfish 150 205.92 203.60 229.44 207.30 217.21 197.30 213.35 213.35 213.35 213.35 225.65 255.62 *Mulloidichthys martinicus* yellow goatfish 160 214.80 212.64 198.06 205.93 213.96 212.21 211.56 223.00 225.21 202.50 197.59 220.19 *Scarus taeniopterus* princess parrotfish 240 265.32 257.42 261.86 266.19 261.18 269.49 283.91 279.91 279.91 279.91 279.91 266.17 *Scarus vetula* queen parrotfish 250 294.53 307.30 304.65 292.55 297.91 286.40 250.00 250.00 250.00 293.94 293.94 284.86 *Sparisoma chrysopterum* redtail parrotfish 240 271.69 269.45 269.09 270.79 265.01 268.37 275.67 278.52 272.03 276.67 273.10 275.68 *Sparisoma viride* stoplight parrotfish 220 285.47 275.90 279.21 279.99 273.11 273.80 275.60 293.78 283.76 280.07 275.25 278.38

Table 2.1: Time-series of Lbar estimates from TIP for all species with sufficient data series.

Table 2.1: Cont.

Table 2.2: Current status (2008-2010) Lbar estimates for all species with sufficient sample data from both the TIP and MRFSS surveys.

Figure 2.1: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP & MRFSS, 2008-2010) for red hind.

Figure 2.2: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP & MRFSS, 2008-2010) for dog snapper.

Figure 2.3: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP, 2008-2010) for schoolmaster.

Figure 2.4: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP, 2008-2010) for bluestriped grunt.

Figure 2.6: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP, 2008-2010) for redtail parrotfish.

Figure 2.7: Length frequency data (TIP, 2008-2010), bounding Lc-Lmax range, and calculated Lbar (TIP & MRFSS, 2008-2010) for redtail parrotfish.

Figure 2.8: Map of Puerto Rico showing fishery-independent sampling sites. Red dots indicate SEAMAP sites and green dots indicate NOAA BioGeo sites.

3.0 THEORY OF LENGTH-BASED STOCK ASSESSMENT

Indicators are needed to assess reef fisheries and to support the implementation of an ecosystem approach to fisheries (Jennings 2005; Cury & Christensen 2005). The principal stock assessment indicator variable we used to quantify population status for the community of Puerto Rican reef fishes was average length (\overline{L}) of the exploited part of the population, which is a metabolic-based indicator that is highly correlated with population size (Beverton & Holt 1957; Ricker 1963; Pauly & Morgan 1987; Ehrhardt & Ault 1992; Kerr & Dickie 2001; Jennings *et al.* 2007). For exploited species, \overline{L} directly reflects the rate of fishing mortality through alterations of the population size structure (Beverton & Holt 1957; Quinn & Deriso 1999). Theoretically, \overline{L} at time *t* is expressed as

$$
\overline{L}(t) = \frac{F(t)\int_{a_c}^{a_{\lambda}} N(a,t) L(a,t) da}{F(t)\int_{a_c}^{a_{\lambda}} N(a,t) da},
$$
\n[1]

where a_c is the minimum age at first capture, a_{λ} the oldest age in the stock, $N(a, t)$ the abundance for age class *a*, *L(a, t)* the length at age *a* and *F(t)* is the instantaneous fishing mortality rate at time *t*. In practice, \overline{L} is usually estimated from lengths in the range of length at first capture *Lc* (or recruitment to the exploited phase of the stock) to the maximum observed length L_{λ} , the length of a fish at a_{λ} . $F(t)$ could also be the viewing power of divers in fisheryindependent visual surveys of reef fish populations (Ault *et al.* 1998).

Using estimates of \overline{L} in time *t*, total instantaneous mortality rate $\hat{Z}(t)$ are estimated using the method of Ehrhardt and Ault (1992)

$$
\left[\frac{L_{\infty}-L_{\lambda}}{L_{\infty}-L_{c}}\right]^{\frac{2(t)}{K}}=\frac{\hat{Z}(t)(L_{c}-\overline{L}(t))+K(L_{\infty}-\overline{L}(t))}{\hat{Z}(t)(L_{\lambda}-\overline{L}(t))+\left(L_{\infty}-\overline{L}(t)\right)},
$$
\n[2]

where K and L_{∞} are parameters of the von Bertalanffy growth equation. Estimates of Z are computed using an iterative numerical algorithm (computer program LBAR; Ault *et al.* 1996; FAO [Food and Agriculture Organization of the United Nations] 2003). An illustrative graph of the relationship between average size in the exploited phase (Lbar) and fishing mortality rate (F) is shown in **Fig. 3.1**.

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Figure 3.1.- Expected progression of average size in the exploited phase dependent of instantaneous fishing mortality rate.

4.0 POPULATION-DYNAMICS DATA & SOURCES

Life history parameters for maximum age, growth and maturity for the reef fish species considered (**Table 4.1**) were obtained from the literature syntheses of Ault *et al.* (1998, 2005*b*) and Claro *et al.* (2001).

Table 4.1: Life-history trait input parameters from literature syntheses of Ault *et al.* (1998, 2005*b*) and Claro *et al.* (2001), and resulting population mortality and sustainability benchmarks estimated from numerical models for exploited Puerto Rican reef fishes.

Example growth and survivorship functions are shown in **Fig. 4.1**. Age-specific functions of length, weight, biomass, and fecundity are shown for red hind in **Fig. 4.2**. Relationships between the spawning potential ratio and the fishing mortality rate at maximum sustainable yield are plotted for a variety of reef-fish species in **Fig. 4.3**. Relationships between maximum size and natural mortality rate and growth rate are shown in **Fig. 4.4**.

Figure 4.1: Growth and mortality dependent on age population demographic relationships for a typical Puerto Rico reef fish.

Figure 4.2: Inter-relationships of growth and mortality dependent on age in the context of population biomass and fecundity (reproductive potential) for red hind, a typical Puerto Rico reef fish.

Figure 4.3: Estimated spawning potential ratio (SPR) at maximum sustainable yield (MSY) dependent on F_{msy} for 25 exploited species (groupers = dark circles, snappers = shaded circles and grunts = open circles) from Puerto Rico. The horizontal dashed line is the 30% SPR USA federal standard for sustainability.

Figure 4.4: Relationships between maximum size (weight in kg) and (*a*) natural mortality rate *M* (y⁻¹), and (b) Brody growth coefficient *K* (y⁻¹) for 25 species (groupers = dark circles, snappers = shaded circles and grunts = open circles) of Puerto Rican coral reef fish.

5.0 NUMERICAL POPULATION MODELING

We used a stochastic length-based numerical population model (Ault & Olson 1996; Ault *et al.* 1998) to calculate ensemble numbers at given lengths \widetilde{N}_{γ} over time for a given cohort γ , generalized as

$$
\widetilde{N}_{\gamma}(L_{\gamma},t) = \int_{a_{\gamma}}^{a_{\lambda}} R_{\gamma}(\tau - a) S(a) \theta(a) P(L \mid a) da ,
$$
\n[3]

where $R_{\gamma}(\tau - a)$ is cohort recruitment lagged back to birth date, *S*(*a*) is survivorship to age *a*, θ *(a)* is a logistic model of sex ratio at age to account for hermaphroditic life histories common to tropical reef fishes, and *P(L|a)* is the probability of being length *L* given the fish is age *a*. This population model simulates the time-transition of recruits to mature adults to maximum sizeage using a number of dynamic functions to regulate population birth, growth, and survivorship processes, including fishery harvests (details in Ault *et al*. 1998). A conceptual flowchart of the REEFS model is shown in **Fig. 5.1**.

REEFS Model Reef Ecosystem Exploited Fishery Simulator

Figure 5.1: Flowchart of the REEFS (Reef Ecosystem Exploited Fishery Simulation) model.

We calibrated the numerical model (Eq. 3) through a consistency check between model estimates of \overline{L} , using \hat{Z} from Eq. (2) as the input, and the $\hat{\overline{L}}$ estimated from data. Additionally, we evaluated the two major components of *Z*, namely fishing mortality rate *F* and natural mortality rate *M*. In this process, we estimated *M* from lifespan applying the procedure of Alagaraja (1984; *sensu* Hoenig 1983) assuming that 5% of a cohort survives to the maximum age/size, and *F* was estimated by subtracting *M* from *Z* (Ault *et al*. 1998). We used the calibrated model to compute management benchmarks of stock status to evaluate sustainability in the following analytical process.
6.0 SUSTAINABILITY BENCHMARKS AND RESOURCE RISK ANALYSES

6.1 Description

Sustainability analyses involved comparison of various population metrics at current levels of fishing mortality against standard fishery management sustainability benchmarks. We configured the simulation model to assess several reference points to address several sustainability risks, including fishery yields, spawning potential ratio (SPR; Clark 1991) and precautionary control rules (for example Restrepo & Powers 1999). Since population biomass *B*(*a*,*t*) is the product of numbers-at-age times weight-at-age *W*(*a*,*t*), yield in weight Y_w from a species during an instant *t* was calculated as

$$
Y_{w}(F, L_{c}, t) = F(t) \int_{L_{c}}^{L_{\lambda}} B(L \mid a, t) dL = F(t) \int_{L_{c}}^{L_{\lambda}} N(L \mid a, t) W(L \mid a, t) dL
$$
 [4]

We obtained an important measure of stock reproductive potential, spawning stock biomass (SSB) at a given level of fishing mortality, by integrating over individuals in the population between the size of sexual maturity (*Lm* ; 50% maturity, assumed knife-edged) and the maximum size (*L*)

$$
SSB(t) = \int_{L_m}^{L_2} B(L \mid a, t) dL \qquad [5]
$$

Maximum spawning biomass is obtained under conditions of no fishing mortality. Spawning potential ratio (SPR) is a management benchmark that measures a stock's potential to produce yields on a sustainable basis, and is computed as the ratio of current *SSB(t)* relative to that of an unexploited stock.

$$
SPR = \frac{SSB_{\text{exploited}}}{SSB_{\text{unexploited}}}
$$
 (6)

Estimated SPRs were compared to USA Federal standards which define 30% SPR as the threshold below which a stock is no longer sustainable at current exploitation levels (see Gabriel *et al.* 1989; Restrepo *et al*. 1998). Evaluation of control rules involved determination of Fmsy (F generating maximum sustainable yield, MSY) and Bmsy (population biomass at MSY) (**Fig. 6.1.1**). We defined $F = M$ as a proxy for F_{msy} (Quinn & Deriso 1999; Restrepo & Powers 1999).

Figure 6.1.1: Conceptual diagram showing limit and target control rules. Target control rules specify desirable levels of fishing for sustainable stocks (for example F(OY) that produces optimal yield OY). Limit control rules define sustainability benchmarks or a cut-off above which there is an unacceptable risk of serious or irreversible harm to the resource and requires strong management intervention. If the maximum fishing mortality threshold (MFMT, equivalent to the F(MSY) limit in our analysis) is exceeded, then management actions in the form of reductions in F (or rebuilding plans) must be implemented to reverse the situation and move the stock to the lower right quadrant $(B/B_{\text{msy}} > 1$ and $F/F_{\text{msy}} < 1$). A more precautionary control rule, as suggested by Restrepo and Powers (1999), is to set the threshold MFMT 'safely below' the MSY limit (for example $F(OY) = MFMT = 0.75 \times F(MSY)$).

6.2 Current population status.

Table 6.2.1: Current status exploitation statistics for all computable species (2008-2010).

Figure 6.2.1: Current status (2008-2010) for all assessable species in the (TIP) database.

Figure 6.2.3: Current status (2008-2010) spawning potential ratio for all species with available data.

6.3 Historic Exploitation & Sustainability Benchmarks of Focus Species

A user-friendly software package, Mortality and Assessment and Stock Simulation Tool (MAST), was developed to facilitate training of DNER personnel in applying sustainability benchmark analyses to Puerto Rico reef-fishes (**Appendix A**). Following are MAST output graphics from length-based sustainability analyses applied to seven focus species: red hind, dog snapper, schoolmaster, bluestriped grunt, lane snapper, redtail parrotfish, and stoplight parrotfish.

6.3.1 Red hind

Figure 6.3.1.1: Yield per recruit in numbers under all conditions of equilibrium F and Lc.

Figure 6.3.1.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.1.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

6.3.2 Dog snapper

Figure 6.3.2.1: Yield per recruit in numbers under all conditions of equilibrium F and Lc. ■ 0-0.1 ■ 0.1-0.2 ■ 0.2-0.3 ■ 0.3-0.4 ■ 0.4-0.5 ■ 0.5-0.6 ■ 0.6-0.7 ■ 0.7-0.8 ■ 0.8-0.9 ■ 0.9-1

Figure 6.3.2.2: Yield per recruit in weight under all conditions of equilibrium F and Lc. \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9

Figure 6.3.2.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.2.4: Spawning potential ratio under all conditions of equilibrium F and Lc. \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9 \blacksquare 0.9-1

Figure 6.3.2.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

6.3.3 Schoolmaster

Figure 6.3.3.1: Yield per recruit in numbers under all conditions of equilibrium F and Lc.■ 0-0.1 ■ 0.1-0.2 ■ 0.2-0.3 ■ 0.3-0.4 ■ 0.4-0.5 ■ 0.5-0.6 ■ 0.6-0.7 ■ 0.7-0.8 ■ 0.8-0.9 ■ 0.9-1

Figure 6.3.3.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.3.4: Spawning potential ratio under all conditions of equilibrium F and Lc.

 \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9 \blacksquare 0.9-1

Figure 6.3.3.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

Figure 6.3.4.1: Yield per recruit in numbers under all conditions of equilibrium F and Lc.

 \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.5-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9

Figure 6.3.4.2: Yield per recruit in weight under all conditions of equilibrium F and Lc. \blacksquare 0.02 \blacksquare 0.02-0.04 \blacksquare 0.04-0.06 \blacksquare 0.06-0.08 \blacksquare 0.08-0.1 \blacksquare 0.1-0.12

Figure 6.3.4.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.4.4: Spawning potential ratio under all conditions of equilibrium F and Lc. \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9 \blacksquare 0.9-1

Figure 6.3.4.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

6.3.5 Lane snapper

Figure 6.3.5.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.5.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

6.3.6 Redtail Parrotfish

 \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9

Figure 6.3.6.2: Yield per recruit in weight under all conditions of equilibrium F and Lc. \blacksquare 0.05 \blacksquare 0.05 0.0.1 \blacksquare 0.1-0.15 \blacksquare 0.15 0.2 \blacksquare 0.2-0.25 \blacksquare 0.25 0.3 \blacksquare 0.3-0.35

Figure 6.3.6.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.6.4: Spawning potential ratio under all conditions of equilibrium F and Lc. \blacksquare 0.0.1 \blacksquare 0.1-0.2 \blacksquare 0.2-0.3 \blacksquare 0.3-0.4 \blacksquare 0.4-0.5 \blacksquare 0.5-0.6 \blacksquare 0.6-0.7 \blacksquare 0.7-0.8 \blacksquare 0.8-0.9 \blacksquare 0.9-1

Figure 6.3.6.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

6.3.7 Stoplight Parrotfish

Figure 6.3.7.3: Eumetric line. Calculates the length at first capture that obtains the maximum yield for every potential value of F.

Figure 6.3.7.4: Spawning potential ratio under all conditions of equilibrium F and Lc.

■ 0.0.1 ■ 0.1-0.2 ■ 0.2-0.3 ■ 0.3-0.4 ■ 0.4-0.5 ■ 0.5-0.6 ■ 0.6-0.7 ■ 0.7-0.8 ■ 0.8-0.9 ■ 0.9-1

Figure 6.3.7.5: Annual exploitation rate with 95% confidence bounds relative to optimum sustainable yield at F=M using (TIP) data from (1988-2010).

7.0 DISCUSSION & FUTURE OPPORTUNITIES – RECOMMENDATIONS

The results of this assessment continue to support the conclusion that the majority of species in the Puerto Rican fishery are significantly over exploited. These populations will benefit greatly from decreased fishing pressure and increased minimum size of exploitation. There does appear to be evidence that this is occurring organically within the commercial fishery as seen in the TIP data. The retirement of older commercial fishers has not been matched by new entries leading to a consistent decrease in effort, in concert with this a shift in gears from trap to hook and line/spear has taken significant pressure off of the smallest individuals through the inherent differences in gear selectivity pattern. However management action is still required as many species are still highly overexploited. Also of major concern is the recreational fishing fleet, which at present has limited impact but possesses the potential for rapid expansion which would negate and possibly reverse the positive trends being seen from the changing commercial fishery. Particular concern should be given to the parrotfish complex which display extreme overexploitation and the grouper complex which while virtually absent still make intermittent appearance in the catch data suggesting that their populations have already undergone collapse but may once have been as prolific as they are at other locations in the Caribbean.

The following figures (**Figs. 7.1-7.3**) will facilitate discussion of applications of length-based assessment methods to Puerto Rico reef-fishes and future collaborative research opportunities.

Figure 7.1: YPR and SPR as a function of F.

Figure 7.2: Comparative spawning potential ratio (SPR) analysis for 32 exploited reef fish species from the Puerto Rican coral reef ecosystem for the period 2008–2010. Red bars indicate overfished stocks, blue bars indicate stocks that are above the 50% SPR, and yellow bars indicate that stocks are between 20-50% SPR.

Figure 7.3: F/F_{msy} ratio versus B/B_{msy} ratio for the 32 species analysed for the years 2008–2010.

ACKNOWLEDGEMETNS

We thank Craig Lilyestrom, Puerto Rico Dept of Natural and Environmental Resources, and Aida Rosario and Daniel Matos, DNER Fisheries Research Laboratory, for their astute technical assistance in obtaining fishery-independent and fishery-dependent databases and for managing logistics of the two sets of regional workshops. We also thank Tony Chatwin, Michelle Pico and Erin Duggan of the National Fish and Wildlife Foundation (NFWF) for their support and encouragement of this project. This research was supported by NFWF Grant #2010-004-2000.

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APPENDIX: MAST Model Description & User's Guide

MANUAL FOR MORTALITY ASSESSMENT AND STOCK SIMULATION TOOL (MAST)

Version 1.0 - August 20th, 2011 Authors: Marc Nadon, Nathan Vaughan, Jerry Ault Programmers: N. Vaughan, M. Nadon University of Miami, Rosenstiel School of Marine and Atmospheric Science

INTRODUCTION

MASTaggregates all analytical tools necessary for the management of exploited fish populations based on length mortality estimates and presents these in a user-friendly visual interface. MAST is coded in JAVA 7.It is composed of 3 general sections: 1) length-based mortality estimation, 2) theoretical models to find optimal fishing regulations, and 3) exploited population simulator. It will also soon have the capability of running stochastic processes through Monte Carlo simulation in order to evaluate uncertainty and risk associated with management scenarios.

Mortality, maturation, and growth-rates are density-independent (on the recruited population).Recruitment can be either density-independent or -dependent (through a stockrecruitment relationship).

The user interface and mortality estimators were coded by N. Vaughan while the population dynamics models and simulation toolwere mainly coded by M. Nadon, all under the supervision and guidance of Dr. Jerry Ault.

1 -GENERAL INFORMATION

TIMING

All time step computations in MAST are in days (i.e. fish ages and simulated times). All yearly parameters (e.g. K, M, etc.) are automatically converted into daily parameters by dividing by 365 (or other conversion steps). Parameters related to transitional management scenarios (e.g. transitional fishing mortality or minimum size-at-first-capture) are also converted into daily time steps. The daily time steps allow for great flexibility in the specification of maturity, recruitment, and mortality temporal patterns (e.g. seasonal closure, periodic recruitment). However, MAST only currently allows inputs in yearly increments.

MORTALITY

Fish cohorts enter all computational matrices at settlement (i.e. transition from pelagic larvae to bottom- or reef-associated fish). The number of recruits can either be specified through a stockrecruitment relationship or can be fixed at a specific value (see *recruitment* section below). Once the initial number of recruits is specified, MASTcalculates initial number in each daily age group using an exponential mortality formula:

$$
N_{t+1} = N_t \cdot e^{-(M+F_t)}
$$

Thesestarting N values can be converted toaverage numbers, when needed,using the following equation:

$$
\overline{N_t} = \frac{N_t}{M+F} \times (1 - e^{-(M+F_t)})
$$

Natural mortality rates are either specified by the user or are derived from estimates of longevity using either a rule-of-thumbs approach that assumes that 1% or 5% of population numbers are left at maximum. This translates into the following general equation:

$$
M = \frac{-\ln(S)}{t_{\lambda}}
$$

where S is the survivorship to age t_A (0.01 or 0.05). A 5% survivorship is a more conservative estimate of natural mortality and is set as the default in MAST.

Fishing mortality *F*is obtained by multiplying the *potential* fishing mortality rate (i.e. the instantaneous mortality coefficient of a fully available age group, which is provided by users) by the gear selectivity *S* which represents the fraction of an age group that is vulnerable to fishing.

$$
F_t = q \cdot f \cdot S_t
$$

Where q is the catchability coefficient (i.e. proportion of stock caught by a single unit of fishing effort), f is the fishing effort, and S is selectivity at age *t*.Fishing mortalities below minimum ageat-first-capture and above maximum observed-age (i.e. t_{λ}) are automatically set to zero (knifeedge selection if selectivity is set to 1 for all age groups).

GROWTH

MAST currently only has the option of using the Von Bertalanffy growth equation to determine length-at-age.

$$
L_t = L_{\infty} (1 - e^{-K(t - t_0)})
$$

Weight-at-age (W_t) is obtained from converting lengths into weights using the equation:

$$
W_t = A \cdot L_t^B
$$

Important: it is critical to use the proper A and B coefficient for specific length and weight units. Most published A and B values convert length in **cm** into weight in **grams**. MAST will convert those weights into kilograms by dividing by 1000 (all weight, biomass outputs are in kg). We therefore highly recommend inputting all length information in cm (e.g. Lint, Lc, Lmax, etc.). It is possible to input length information in other units (e.g. mm) if the proper A and B coefficient are used (i.e. those that will output weight in grams).

It is also critical to use the same measurement of fish length for all inputted data (e.g. fork, standard, or total length).

MATURITY

Users can enter either age- or size-at-maturity (MAST will automatically convert one to the other). Maturity is currently set to be knife-edged (i.e. 100% of a cohort reaches maturity at a specific age or size). A logistic curve may be available in future versions, if the need arises.

RECRUITMENT

For simplicity, recruitment levels are often set at a fix value and assumed to be densityindependent, especially if spawning stock biomass is known to be at a safe level (SPR > 30%).Alternatively, recruitment can be dependent on spawning stock biomass (SSB).Current recruitment (R_t) is generally assume to be a function of the spawning stock biomass (SSB) at a certain time in the past equivalent to hatching time + pelagic larval stage duration (d).

$$
R_t = SSB_{t-d} \cdot f(SSB_{t-d})
$$

The Beverton-Holt version of this general equation (Ricker's equation will be available soon) is:

$$
R = \frac{SSB}{\alpha + \beta \cdot SSB}
$$

Using Francis (1992)'s re-parameterization,the two parameters of this equation can be defined as

$$
\alpha = \frac{B_0(1-h)}{4hR_0} \quad \text{and} \quad \beta = \frac{5h-1}{4hR_0}
$$

whereB₀ is an estimate of pristine spawning stock biomass, R₀ is number of recruits at B₀,*h*is the steepness of the initial stock-recruitment curve (i.e. fraction of R_0 corresponding to spawning

stock biomass at 20% B₀). To use this stock-recruitment relationship, users need to define both B₀ and *h*. They also need to input the larval duration in days (i.e. time between spawning event and settlement). Spawning schedule is currently set to be continuous throughout the year, but extra functionality will be added later, with the capability to define recruitment seasonality more precisely.

Below are examples of Beverton-Holt stock-recruitment curves with varying steepness *h*.

2 - LENGTH-BASED MORTALITY ESTIMATION

BEVERTON-HOLT MODEL

Beverton and Holt (1954) were the first to derive an equation relating average length in the catch (Lbar) to total mortality (Z) .

$$
\overline{L} = \frac{\int_{t_c}^{t_{\infty}} F_t \cdot L_t \cdot N_t \cdot dt}{\int_{t_c}^{t_{\infty}} F_t \cdot N_t \cdot dt} \rightarrow Z = \frac{K \cdot (L_{\infty} - \overline{L})}{(\overline{L} - L_c)}
$$

However, Ehrhardt and Ault (1992) found this model to be biased because of the integration to infinite age (i.e. influenced by theoretical, very old, fish that are never present in catch records). This model is not available in MAST due to this problem.

AULT-EHRHARDT MODEL

Ault and Ehrhardt (1991) proposed a truncated version of the Beverton-Holt model that sets a realistic upper limit for maximum lengths (L_{max} or L_{λ}). As such, this model takes an extra parameter (t_λ or L_λ).

$$
\overline{L} = \frac{\int_{t_c}^{t_{\lambda}} F_t \cdot L_t \cdot N_t \cdot dt}{\int_{t_c}^{t_{\lambda}} F_t \cdot N_t \cdot dt} \rightarrow \left(\frac{L_{\infty} - L_{\lambda}}{L_{\infty} - L_c}\right)^{Z/k} = \frac{Z(L_c - \overline{L}) + K(L_{\infty} - \overline{L})}{Z(L_{\lambda} - \overline{L}) + K(L_{\infty} - \overline{L})}
$$

This model does not have the same bias as the Beverton-Holt model, but, as with the B-H model, it does assume equilibrium conditions (i.e. stable recruitment and mortalities during a time period long enough for stock age structure to be stable).

VAUGHAN-AULT MODEL

An improved model is currently under development that will be able to deal with nonequilibrium mortality conditions. This model will merge the size-truncated model of Ehrhardt and Ault (1992) with the non-equilibrium model develop by Gedemke-Hoenig (which is based on the Beverton-Holt model and thus suffers from the same potential bias). This tool is not currently available in MAST 1.0.

3 - THEORETICAL MODELS USED TO ESTIMATE OPTIMAL MANAGEMENT REGULATIONS

Once current fishing mortality rates are estimated, it is possible to parameterize various models in order to estimate current stock status and preferable management targets. If recruitment is set at a fix level, users can run yield-per-recruit (YPR) and spawning potential ratio (SPR) analyses. If a stock-recruitment function is defined, users can run models in terms of absolute yield, which take the effect of reduced spawning stock biomass on recruitment (and yield) into account. These models all assume that a population has reached equilibrium.

YIELD-PER-RECRUIT

MAST calculates YPR in "piece-wise" fashion by applying the mortality equations defined above to a fixed number of recruits (e.g. 1000) all the way to maximum age (t_{λ}) , using daily increments. YPR is calculated at each daily age by multiplying average biomass by fishing mortality F, summing all daily yields, and dividing by the original number of recruits.

$$
YPR = \frac{1}{Recruits} \sum_{t=0}^{t_{\lambda}} F_t \cdot \overline{N}_t \cdot \overline{W}_t \cdot
$$

YPR is calculated for a large number of combinations of length-at-first-capture (L_c) and fishing mortality rates (F). Specifically, YPRsare calculated from $L_c = 1$ cm to L_{max} in $L_{\text{max}}/100$ increments, and from $F = 0$ to 2.5 in increments of 0.025. The YPRs for all these combinations are exported from MAST and can be plotted in Excel or other software (see examples at the end of this section).

SPR

MAST calculates spawning potential ratio (SPR) in a similar way as for YPR. Biomass is calculated at each daily age and spawning stock biomass is simply the sum of all biomass above the age at maturity.Spawning stock biomass is calculated as

$$
SSB = \sum_{t=t_m}^{t_{\lambda}} \overline{N}_t \cdot \overline{W}_t
$$

where W_t is the average weight-at-age. Spawning stock biomass is calculatedat pristine levels (Bo) and under various management scenarios (L_c, F) , similarly to YPR. SPR is then calculated for different levels of L_c and F by dividing spawning stock biomass under exploitation by pristine spawning stock biomass.

$$
SPR = \frac{SSB_F}{SSB_{F=0}}
$$

SPR isopleth graphs can be produced from the MAST output, similarly to YPR isopleth (see example below).

ABSOLUTE YIELD WITH VARIABLE RECRUITMENT

If a recruitment function is defined, yield can be calculated in absolute terms. MAST first calculates the number of recruits entering the population under equilibrium conditions (i.e. for specific L_c and F values). The following equation is used

$$
R_e = \frac{SSB_e - \alpha}{\beta \times SSB_e}
$$

whereα and $β$ are parameters of the Beverton-Holt stock-recruitment equation. From the equilibrium recruitment level, the structure of the population at equilibrium is derived from which absolute yield can be calculated.As for YPR and SPR, absolute yield is calculated for various combinations of L_c and F and can be plotted in Excel.

Below are examples of SPR (left) and absolute yield (right) graphs created in SigmaPlot.

4 - EXPLOITED POPULATION SIMULATION

MAST includes population simulation capabilities which allow users to track Lbar, yield, or SPR forward through time in daily time steps according to various management scenarios (i.e. L^c and/or F changing at various times). To use this tool, users need to provide vectors of expected 'future' yearly changes in F and Lc.

The simulation can be run with fixed recruitment, set at a specified level. It can also be run with a stock-recruitment function after specifying pristine spawning stock biomass (B_0) , steepness (h) , and pelagic larvae duration.

Note: MAST is capable of running population simulations forward through time for a (theoretically) unlimited number of daily time steps, given some basic computer memory capacity (i.e. at least 256 mb). MAST determines available memory to JAVA and divides the simulation task in manageable blocks. The more memory a computer has, the larger the time blocks and the faster the simulation will run through completion.

Below is an example of a MAST simulation output based on Mutton snapper life history (L_c = 300 m

Sustaining Coral Reef Fisheries of Puerto Rico

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> Report of Technical Workshop #3 May 2014

ROSENSTIEL SCHOOL of MARINE & ATMOSPHERIC SCIENCE

Conveners: Drs. Jerry Ault and Steve Smith, University of Miami's Rosenstiel School of Marine and Atmospheric Science.

Local Host: Dr. Craig Lilyestrom, Puerto Rico Dept. of Natural and Environmental **Resources**

Location: DNER, San Juan: May 19-20, 2014.

Introduction.- The coral reef fisheries of the Puerto Rico reef ecosystem support multimillion-dollar fishing and tourism industries. The sustainability of these fisheries is a key conservation concern given their economic and ecological importance, the significant dependence of subsistence and artisanal fishers on reef fisheries for their livelihoods, and the considerable and growing threats to coral reef habitats (i.e. coral bleaching and disease, pollution and climate change). Sustainability refers to the ability of an exploited stock to produce goods and services, including yields at suitable levels in the short term, while maintaining sufficient stock reproductive capacity to continue providing these goods and services into the indefinite future. The data- and modellimited situations confronting most coral reef fisheries, including those of Puerto Rico, have hampered application of modern stock assessment techniques that meet the legal mandate of the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA). Technical Workshops #1 and #2, conducted in May and September 2013, focused on the theory and application of length-based methods for stock assessment of data-limited fisheries in Puerto Rico. These methods were applied to 34 reef-associated species that had sufficient length composition and life history data to assess current sustainability status. Additionally, in-depth assessments of changes in sustainability status over the past 23 years were conducted for seven focus species: red hind, dog snapper, schoolmaster, bluestriped grunt, lane snapper, redtail parrotfish, and stoplight parrotfish. Collaborating scientists and managers from the Puerto Rico Department of Natural and Environmental Resources (DNER) subsequently developed a suite of fishing regulations, including gear and size restrictions, to correct overfishing problems identified in Technical Workshops #1 and #2. These regulations are pending final legal approval.

Another outcome of Technical Workshops #1 and #2 was the identification of critical data needs to support future assessment and management of coral reef fisheries in Puerto Rico. Addressing these data needs became the focus of Technical Workshop #3.

Workshop Goal:- The goal of Technical Workshop #3 was reflected in its title, "Building Fisheries Information Systems for Sustaining Coral Reef Fisheries of Puerto Rico".

Workshop Objectives:- The objectives of Technical Workshop #3 were twofold: (1) to create a more accessible database for information on recreational fishing in Puerto Rico; and (2) to prioritize key additional data needs and to discuss strategies for addressing these needs over the next several years. Accomplishments for these two objectives are described in the following sections.

Accessible Database for Recreational Fishing:- Collaborating researchers from UM and DNER developed data processing procedures and associated computer code for creating a more accessible database for the Marine Recreational Fisheries Statistical Survey (MRFSS) in Puerto Rico, the primary sampling program for the recreational fishing fleet which began in 2000. The intercept survey form for recreational fishers in Puerto Rico is provided in the Appendix. A relational database for MRFSS data for Puerto Rico was designed using Microsoft Access. Researchers at the University of Miami (UM) and DNER collaborated on compiling the various tables describing codes for key variables such as species, Municipios, site locations, fishing gears, etc. UM researchers developed data processing procedures using SAS statistical software to transform the NOAA MRFSS data into the various key data tables for Puerto Rico. The overall processing flow is illustrated in the diagram of **Fig. 1.** The three primary relational data tables for trip, catch, and length-weight information are respectively described in **Tables 1**, **2** , and **3**.

Figure 1. Processing flow diagram for transforming NOAA's database for the Marine Recreational Fisheries Statistical Survey for Puerto Rico into a user-friendly relational database in Microsoft Access. The processing procedures utilize SAS statistical software and Microsoft Excel.

Table 1. Description of variables in the Trip information table of the MRFSS Access database for Puerto Rico .

Table 2. Description of variables in the Catch information table of the MRFSS Access database for Puerto Rico .

Table 3. Description of variables in the Length information table of the MRFSS Access database for Puerto Rico .

Key Data Needs for the Next Several Years:- Discussions of key data needs for the near-term future for building sustainable fisheries in Puerto Rico began with a review of fishing regulations pending final legal approval for correcting some of the overfishing problems identified in Technical Workshops #1 and #2. To date, sustainability analyses have been conducted on 34 reef-associated fish species; however, there are a number of reef species that are prominent in the landings of commercial and recreational fishers and thus have sufficient length composition data for sustainability analysis, but are lacking the requisite life history information on age & growth, lifespan, length at reproductive maturity, etc. These include species of the following taxa groups: deepwater snappers, parrotfishes, boxfishes, porgies, and goatfishes. Workshop participants unanimously agreed that conducting research on life history characteristics of these species should be a high priority in the next few years. Participants sketched out a blueprint for future research and funding to improve DNER's capacity for conducting life history studies.

Technical Workshop #3 concluded with a review and discussion of the principal sources of length composition and abundance data for Puerto Rico reef-fishes. Great strides have been made in the past few years in synthesizing and assimilating data from the Trip Interview Program (TIP) that samples the commercial fleet. Researchers from UM are currently working with their counterparts at NOAA's Southeast Fisheries Science Center to create a master database for Puerto Rico TIP information. This activity involves cross-checking and updating data records for the period 1988-2010 utilizing the original source digital files provided by DNER. Workshop participants were also briefed on a new fishery-independent diver visual survey of Puerto Rico reeffishes to be conducted during the summer of 2014. Participants sketched out a blueprint for future research and funding to integrate fisheries databases for the commercial fleet (Trip Interview Program), the recreational fleet (MRFSS), and fisheryindependent surveys.

Appendix MRFSS Intercept Data Recording Form

